

NITROGEN FIXATION AND GROWTH RESPONSE OF *Alnus rubra* AMENDED WITH LOW AND HIGH METAL CONTENT BIOSOLIDS

Linda S. Gaulke^{1*}; Charles L. Henry²; Sally L. Brown³

¹ Civil and Environmental Engineering, University of Washington - 333 More Hall/Box 352700 - Seattle, WA - 98195-2700 - USA.

² Interdisciplinary Arts and Sciences, University of Washington, Bothell Campus - USA.

³ College of Forest Resources, University of Washington - Box 352100 - Seattle, WA - 98195 - USA.

*Corresponding author <lsg@u.washington.edu>

ABSTRACT: Forest application of biosolids offers a potential environmentally friendly alternative to landfilling. This two-year investigation was designed to analyze the effects of elevated soil metal concentration resulting from the land application of biosolids on the symbiotic, nitrogen (N) fixing relationship between *Alnus rubra* Bong. (red alder) and *Frankia*. High metal biosolids and a modern-day composted biosolid applied at high loading rates of 250, 500, and 1000 Mg ha⁻¹, were used to represent a worst-case scenario for metal contamination. The high metal biosolids were obtained before the current regulations were formulated and had been lagooned prior to use in this study. Total cadmium (Cd), lead (Pb) and zinc (Zn) in the high metal biosolids were 45, 958, and 2623 mg kg⁻¹ respectively. These metal concentrations are above current regulatory limits in the US. The compost was made using biosolids that are currently produced and had Cd, Pb and Zn of 0.8, 20 and 160 mg kg⁻¹ respectively. Trees were harvested and analyzed for rate of N fixation (as measured by acetylene reduction activity), biomass, and foliar metals. Soils were analyzed for available N, total carbon and N, pH and total Cd, Pb and Zn. Rates of N fixation were not affected by soil amendment. In year 2, shoot biomass of trees grown in both the compost and high metal amendments were higher than the control. Shoot biomass increased with increasing amount of compost amendments, but decreased with increasing amount of high metal amendments. There was no relationship between soil metal concentration and plant biomass. Foliar Cd and Pb were below detection for all trees and foliar Zn increased with increasing amount of both compost and high metal amendment, with concentrations of 249 mg kg⁻¹ for trees grown in the compost amendment and 279 mg kg⁻¹ for the high metal amendment. The results from this study indicate that the growth of *A. rubra* benefited from both types of biosolids used in the study and that the *Alnus/Frankia* relationship was not negatively impacted by metal concentrations resulting from the high metal biosolids amendments. Key words: *Frankia*, cadmium, lead, zinc

CRESCIMENTO E FIXAÇÃO DE NITROGÊNIO POR *Alnus rubra* CULTIVADO SOB FERTILIZAÇÃO COM BIOSÓLIDOS COM ALTOS E BAIXOS TEORES DE METAIS

RESUMO: A aplicação de biosólidos em florestas representa uma potencial alternativa ambientalmente correta para aterros sanitários. Os efeitos de fertilização de nitrogênio na relação simbiótica e de fixação de nitrogênio *Alnus rubra* Bong. (amei-ro vermelho) e *Frankia* foram avaliados em um estudo de campo conduzido por dois anos. Um biosólido com concentração de metais muito acima dos marcos regulatórios (45 mg kg⁻¹ Cd; 958 mg kg⁻¹ Pb; 2623 mg kg⁻¹ Zn) produzido há mais de 25 anos e um biosólido composto recente foram aplicados nas doses de 250, 500 e 1000 Mg ha⁻¹, representando um cenário pior possível de concentração de metais. Ao final do segundo ano, a biomassa da parte aérea das plantas cultivadas sob ambos os compostos era maior que a das plantas controle, mas não foi observada inter-relação entre a concentração de metais no solo e a biomassa das plantas. As concentrações de Cd e Pb foliares permaneceram abaixo do nível de detecção em todas as plantas, mas a concentração de Zn foliar aumentou proporcionalmente ao aumento da adição do bioremediador, atingindo concentrações de 249 mg kg⁻¹ e 279 mg kg⁻¹ para os tratamentos alto metal e composto, respectivamente. Embora não tenham sido registrados efeitos negativos do biosólido com altos teores de metais, outras rotas de exposição podem levar a possíveis impactos ecológicos.

Palavras-chave: *Frankia*, cádmio, chumbo, zinco

INTRODUCTION

As a result of pre-treatment regulations, metal concentrations in biosolids in the US have been greatly reduced over historic levels. As wastewater treatment programs outside of the US become more common, questions will arise on appropriate metal limits for land application of these materials. There are still uncertainties around the potential for the metals in biosolids to have a negative effect on soil ecosystems (Horswell et al., 2003). The bioavailability of metals to soil microorganisms is a factor to be considered in this evaluation. For example, in the UK limits for total zinc (Zn) in biosolids were set to protect *Rhizobium leguminosarum* biovar *trifolii* (Giller et al., 1999; McGrath & Chaudri, 1999). This action was taken because *R. leguminosarum* was considered an indicator organism (Smith, 1997). Results from a long-term study at the Woburn Market Garden using repeated, high-loading applications of biosolids showed that *R. leguminosarum* had lost its ability to fix nitrogen (N) (Chander & Brookes, 1993). This decline was attributed to soil Zn concentrations. These results are not definitive however, as other long-term studies looking at *R. leguminosarum* have not seen the same effect (Giller et al., 1998; Obbard & Jones, 1993; Smith, 1997).

In forestry, an analogous relationship is formed by the *Frankiaceae*. *Frankia* are a type of filamentous actinomycete that forms a symbiotic N fixing relationship with actinorhizal plants (e.g. *Alnus*). Although both *Alnus* (Kramer et al. 2000a; b) and *Frankia* (Richards et al., 2002; Sayed, 2003; Sayed et al., 2000) have exhibited signs of metal tolerance, the effect of soil metals on *Frankia* is not fully understood. As *R. leguminosarum* has been used as an indicator of microbial health in agricultural systems, N fixation by *Frankia* may provide information for microbial health in forest ecosystems.

This study was conducted to evaluate the impact of high cadmium (Cd), lead (Pb), and Zn concentrations resulting from the land application of biosolids on the growth of *A. rubra* as well as on the N fixation of the *A. rubra*-*Frankia* relationship.

MATERIALS AND METHODS

Study Site Description

This study was conducted at a research forest located in Eatonville, Washington, US (47°37'N, 122°19'W) at the base of the foothills of the Cascade Mountains. Average annual rainfall is 1200 mm, with the rainfall from July and August typically being less than 120 mm (Pack Forest precipitation records). The study site is 317m above sea level and its soil is a well-

drained, sandy, glacial outwash soil in the Indianola series (mixed, mesic Dystric Xeropsamments) (NRCS, 2004).

Site Installation

Prior to installation, which took place in February 2002, the site was densely vegetated with *Cytisus scoparius* (Scotch Broom) and *A. rubra*. It was cleared of vegetation and the soil was disked to break up any remaining roots for removal. Two biosolid amendments from the same municipality (King County, Washington) were used in this study, 25 year-old biosolids and composted biosolids, along with a control. The 25 year-old biosolids had been lagooned prior to the start of this study and were used to represent high metal biosolids (hereafter referred to as high metal). They were produced in King County using an anaerobic stabilization process prior to the US Environmental Protection Agency (USEPA) code of federal registry (CFR) Part 503 regulations that set limits on metal concentrations in biosolids for land application (USEPA, 1999). The composted material was produced using current anaerobically stabilized biosolids mixed with sawdust (hereafter referred to as compost). The compost was included as stable biosolids with total metal concentrations representative of the average biosolids currently produced in the US. Total Cd, Pb and Zn in the high metal biosolids were 45, 958, and 2623 mg kg⁻¹, respectively. Biosolids that are currently produced in this wastewater treatment plant and that were used to make the compost used in this study have total Cd, Pb and Zn concentrations of 5, 52 and 825 mg kg⁻¹, respectively. High loading amounts of biosolids were used in this study to represent a worst case scenario for metal loadings; 1000 Mg ha⁻¹, 500 Mg ha⁻¹ and 250 Mg ha⁻¹ (dry weight basis) of each type of biosolids. The high metal biosolids were used for research purposes only and would not be permitted to be land applied under current US regulations.

This study was designed as a randomized complete block with three replicates of each treatment. Plots were each 2 m × 2 m. There was a 1 m space between each plot as well as a 3 m work row in between each replicate.

Amendments were applied evenly to the surface of each plot and then rototilled to a depth of 15-20 cm. Nine bare-root *A. rubra* Bong. seedlings, 40-50 cm in height, were planted in each plot evenly spaced in a square grid pattern with a 0.5 m buffer between the outermost trees and the edge of the plot. This resulted in a minimum of a 2 m space between trees planted in plots with different treatments within the same replicate. There were 27 trees for each treatment (9 trees per plot, 3 plots per treatment). The trees

were purchased from a local nursery. They had been raised from seeds collected in lowlands in Whatcom and Skagit Counties. At the time of planting, the whips had sparse and small nodules and no leaves.

Maintenance

To eliminate any additional competition for soil available N weeds within the plot area were treated by direct application of glyphosate herbicide to foliage with a wick applicator. The area surrounding plots was mowed to discourage additional weeds. Deer browsing of *A. rubra* foliage was initially controlled by biocontrol methods. In December 2002, a 2.5 m chain link fence was built surrounding the entire study and this eliminated access for deer.

Soil/Biosolids Sampling

Soil samples were collected prior to amendment application, as well as during tree harvest in September 2002 and August 2003. Random grab samples of compost and high metal soil amendments were also collected for analysis prior to installation. Soil samples were collected with a stainless steel probe (2.5 cm diameter) in the 0-15 cm depth range. Three samples were taken from each plot and homogenized. Samples were kept on ice or refrigerated until return to the laboratory when the available N analysis was performed. Available N measurements were conducted on field moist samples; all other soil samples were air-dried for remaining analyses.

Available Nitrogen

Available N extracts were run on unsieved soil samples in order to prevent removal of the woodchips that compose a large portion of the compost amendment. Soil samples were hand picked to remove any visible rocks (Keeney & Nelson, 1982; Robertson et al., 1999). A 2 M potassium chloride (KCl) extraction was used to determine available N (NO_2^- , NO_3^- and NH_4^+) (Keeney & Nelson, 1982; Robertson et al., 1999). Five g of soil were placed in a glass flask with 50 mL 2M KCl. The flasks were placed on a shaker for one hour, left to settle, and then filtered through Whatman No. 41 filter paper. The solution was analyzed with a Lachat flow injection analyzer using QuickChem method 107-04-1-B for $\text{NO}_3^-/\text{NO}_2^-$, and 10-107-06-1-F for NH_4^+ . Ammonium nitrate (NH_4NO_3) solution standards were used to calibrate for concentrations of NO_3^- and NH_4^+ from 0 to 1,000 $\mu\text{g kg}^{-1}$ (calibration curve coefficient = 0.99). Soil water content was used to correct for percent moisture of field moist samples.

pH

Soils slurries were mixed 1:1 on a volume basis with deionized water and air-dried soil samples and

then left to stabilize. The pH was measured after one hour with an Orion pH meter (Bremner, 1996; Nelson & Sommers, 1996).

Total Carbon and Nitrogen

Soil samples were ground with a mortar and pestle and placed into foil packages and weighed on a CAHN C-33 microbalance. Total C and N were determined using a CE-440 elemental analyzer, which combusts samples at 980°C (Bremner, 1996; Nelson & Sommers, 1996). National Institute of Standards and Technology (NIST) standards “Montana soil” and “sludge standard” were used, along with acetanilide and benzoic acid as internal standards. Respective rates of recovery were for standards were: 0.89-0.96 (C, Montana soil), 0.95-1.06 (N, sludge standard), 0.94-1.01 (C, acetanilide), 0.94-0.96 (N, acetanilide), 0.98 (C, benzoic acid). Total C and N were reported as % on a dry weight basis, and the C:N ratio as weight:weight.

Total Metals

Total Cd, Pb and Zn concentrations were analyzed for both types of amendment and soil samples taken prior to installation, and on soil samples collected during the second year of field work. Phosphorus (P) was also measured on soil samples. Soil samples were air-dried and sieved to 2mm. A composite 2 g sample was used for each plot. USEPA method 3050B, acid digestion of sediments, sludges, and soils (USEPA, 1986) was followed for the analysis. Samples were digested and brought to near dryness with concentrated HNO_3 and HCl acids and 30% H_2O_2 . Samples were refluxed and then filtered through Whatman No.41 filter paper and brought to volume with DI water. NIST standards “Montana soil” and “sludge standard” were used, along with an internal solution standard consisting of 0.1 mg L^{-1} each; Cd, Pb and Zn. Respective rates of recoveries for standards were: Montana soil 0.96 Cd, 0.78 Pb, 1.01 Zn, sludge standard 0.79 Cd, 0.83 Pb, 0.93 Zn, and internal solution standard 1.00 Cd, 0.99 Pb, and 1.06 Zn. Samples were analyzed on a Perkin Elmer Optima 3300DV inductively coupled plasma optical emission spectrometer (ICP-OES) at a 1:10 dilution. Solutions were also rerun for Cd on a Perkin Elmer AAnalyst 300 atomic absorption spectrometer (FLAA), using flame with no dilution, to confirm the accuracy of the detection limits that were seen with the ICP-OES.

Tree Harvest

In certain plots, several trees did not survive. The cause of mortality was investigated and determined to be due to either handling of the plants prior to planting, or a hard frost following planting (Edmonds, personal communication, 2002). Two trees from each

plot were harvested for analysis during the first week of September in 2002, and the last week of August in 2003. In certain cases in year 1, one tree was harvested per plot to assure adequate biomass for harvest in year 2. In one plot of the compost 1000 t ha⁻¹ treatment no trees were harvested in year 1. As is described in the results section, the mortalities were not related to tree yield or metal concentration. Missing values for this treatment was estimated as outlined in Zar (1996), including a correction for the sum of squares bias. For the second harvest, two trees were harvested from all plots in all treatments. For both harvests, results from each tree were averaged and the average was used as a single value from each plot for statistical analysis. At the end of year one, trees were thinned in each plot so that a maximum of 4 trees remained for a treatment total of 12 trees. This was done to reduce competition and shading effects.

Nitrogen Fixation

Relative rates of N fixation were measured as acetylene reduction activity (ARA) and reported as micromoles C₂H₄ produced per gram of nodule used in the assay (dry weight basis) per hour (μmol C₂H₄ g⁻¹ h⁻¹) as described in Hardy et al. (1973). Two trees from each plot were harvested for analyses during the first week of September in 2002, and the last week of August in 2003. Values for both trees were averaged and a single value per plot was used for statistical analysis. This time of year was selected for tree harvesting to allow for a reasonably complete growing season for comparison of biomass. In addition, it was determined based on reported midday observations of *A. rubra* ARA in July (34.9 μmol g⁻¹ hr⁻¹), September (20.9 μmol g⁻¹ hr⁻¹) and January (0.1 μmol g⁻¹ hr⁻¹) (Teklehaimanot & Martin, 1999) that ARA would be sufficiently high at this time. Blocks were analyzed on consecutive days at the same time of day to account for diurnal variations. Tree roots were dug up with a garden fork and soil was lightly brushed off. A random selection of nodules was removed from roots with a razor knife, keeping a portion of the root intact to ensure that damage did not occur to nodules, and placed into a 65 mL glass vial, which was then sealed with rubber septa. 6.5 mL air was removed from the vial and then 6.5 mL acetylene (C₂H₂) was added (10% by volume), and the head space was mixed well. Positive controls (nodules, no C₂H₂) and negative controls (C₂H₂, no nodules) were included. The reaction was left to run for three hours in the dark and then a 3ml sample was taken from each vial and stored in a vacutainer for transport to the laboratory. Samples were analyzed for ethylene (C₂H₄) on a SRI 8610C gas chromatograph (GC) with a 60mx0.53 Supelco capillary column using

manual injection with nitrogen as the carrier gas (Hardy et al., 1973; Myrold et al., 1999; Weaver & Danso, 1994). The detection limit for this analysis taking dilution into account was 1.5 micromoles g hr⁻¹.

Biomass

Whole trees were harvested for the determination of biomass. Shoots and roots were separated at the soil line for root and shoot biomass measurements in 2002. All nodules were brushed to remove soil and removed from roots for separate biomass evaluations. Plants were oven-dried for 72 hours at 70°C. Biomass was reported as dry weight in grams. Two trees were harvested from each plot the first year of the study and results were averaged. This average was used as the value for each plot for statistical analysis. For the second year only shoots were harvested as dry cemented soil conditions made it difficult to accurately collect the entire root and nodule system for determination of total biomass. Mean ± standard deviation is reported in the text.

Foliar total metals

Fifteen leaves were randomly collected from each harvested tree the second year, and combined as one sample for each plot. Leaves were dried at 70°C for 72 hours to determine dry weight. Foliar total metals were determined using a laboratory adaptation of the Association of Analytical Communities (AOAC) method 3.014(a) ((AOAC, 1984) as described in Brown et al. (2004). Two g of each composite leaf sample were ashed in glass beakers at 480°C for 16 h. Ashed samples were dissolved in 2 mL HNO₃, placed on a hotplate and heated until near dryness. Samples were then refluxed with 10 mL 3N HCl. Samples were brought to a final volume of 25 mL with DI H₂O and centrifuged. NIST standards “orchard leaves” and “spinach leaves” were used, along with an internal solution standard consisting of 0.1 mg L⁻¹ each; Cd, Pb and Zn. Respective recovery rates for standards were; orchard leaves 0.89 Pb and 1.12 Zn, spinach leaves 0.84 Cd, 100 Pb, 0.98 Zn; and internal solution standard 1.00 Cd, 0.99 Pb, and 1.06 Zn. Samples were analyzed on a Perkin Elmer Optima 3300DV ICP-OES. Cadmium was also analyzed on Perkin Elmer Analyst 300 (AA) with flame to verify detection limits that were seen with the ICP-OES.

Statistical Analysis

Statistical analyses were performed using SPSS 12 for Windows (SPSS Inc., 2003). Data was checked for normality by visual inspection and confirmed using Kolmogorov-Smirnov tests. Analysis of variance (ANOVA) was used to test for main effects of treatment and treatment versus rate interactions (*P*

= 0.05). Means were separated using Duncan's multiple range test ($P = 0.05$). Linear regression was used to evaluate relationships between treatment rate and measured properties, and correlation was evaluated among measured properties ($P = 0.05$).

RESULTS AND DISCUSSION

Soil Properties

Results of total soil Cd, Pb and Zn for year 2 are presented in Table 1. The value for the control soils (1.9 mg kg^{-1}) was relatively high as compared to reported average values in the Pacific Northwest ($\sim 0.30 \text{ mg kg}^{-1}$), but were within the range of naturally occurring soil Cd (Traina, 1999). Metal concentrations were similar for the control and compost treatments and increased with increased loading of high metal biosolids. The metal concentrations in the high metal treatments were at or above those at which negative impacts on soil microbial function have been observed. For example, at total soil Cd 12.4 mg kg^{-1} and Zn 435 mg kg^{-1} , Giller et al. (1989) found *Rhizobium* isolated from nodules of white clover had lost their ability to symbiotically fix N.

Soil NO_3^- , NH_4^+ , total C, N, C:N and pH are presented in Table 2. Available N (NO_3^- , NH_4^+) did not increase with increasing application rates in either the compost or high metal amendments the first year. This may have been a result of the high loading applications used, with a high percentage of the surface soil consisting of amendment. However, for the second year, both soil NO_3^- and NH_4^+ increased with increasing applications of both amendments. Total C and N, along with the C:N ratio, decreased from year 1 to year 2 in all of the compost treatments, but in all other treatments was similar from year 1 to year 2. The C:N ratio (9:1) of the high metal biosolids was within the range commonly seen prior to land application, indicating that the material had not decomposed over time. The slow rate of decomposition of these materials potentially resulted from their storage in an anaerobic en-

vironment for 25 years prior to this study (Henry et al., 1999). In this study, biosolids applied at 13.4 t ha^{-1} supplied plant N equivalent to 140 kg ha^{-1} urea. Both foliar N and tree growth in the biosolid treatments are evidence that this rate was sufficient to meet the N needs of the trees. For this study, total N provided by all plots receiving either compost or high metal biosolids was well in excess indicating that there was sufficient N for plant growth.

Soil pH in the high metal treatments (pH 4.1-4.7) was lower than the control (pH 5.1), with the second year high metal 1000 t ha^{-1} treatment having the lowest pH of all treatments (pH 4.1). Low pH has been found to be an important regulating factor for N fixation in *Rhizobium*. Ibekwe et al. (1997) grew *R. leguminosarum* bv. *trifolii* in high and low metal biosolids amended soils under varying pH conditions (pH range 4.4 to 6.4) and found that N fixation as ARA was greater in all high pH treatments as compared with lower ones. In addition, the number of *Rhizobium* was reduced in all of the low pH soils, including the controls, resulting in reduced and ineffective nodulation. The concentrations for Cd ($\cong 1.0$ and 4 mg kg^{-1}) and Zn ($\cong 133$ and 140 mg kg^{-1}) however, were lower than those in the high metal biosolid treatments in this study. It is not known what the optimum pH range for *Alnus* and *Frankia* nodulation and fixation is. Crannell et al. (1994) provide an exhaustive review of studies reporting *Alnus* and *Frankia* association at a range of soil pH. They report that *Alnus* have been shown to nodulate in soils below pH 4, and that one study showed no difference in forest soil nodulation from pH 4.5 to 7.3. The USEPA CFR part 503 does not have a recommendation for pH, other than that it be monitored and adjusted over time to ensure proper pH for ideal crop growth (USEPA, 1993). Based on previous studies, it would be expected that some negative impact on *Frankia* would be observed as a result of both the metal concentrations and low pH in the high metal treatments in this study.

Table 1 - Total soil Cd, Pb and Zn concentrations for low metal (compost) and high metal biosolid amended soils. Means and standard deviations are presented.

Treatment	Rate	Soil Cd	Soil Pb	Soil Zn
	Mg ha^{-1}	----- mg kg^{-1} -----		
Control		1.2 ± 0.077	28.3 ± 15.2	109.5 ± 52
Compost	250	3.4 ± 1.43	39.1 ± 20.2	161.8 ± 56
Compost	500	3.7 ± 0.32	42.4 ± 5.3	254.4 ± 29
Compost	1000	4.0 ± 0.41	50.2 ± 3.3	283.2 ± 55
High Metal	250	14.8 ± 3.71	258.6 ± 77.5	627.1 ± 198
High Metal	500	19.9 ± 1.05	352.0 ± 11.7	1012.0 ± 16
High Metal	1000	32.5 ± 4.08	623.0 ± 100	1759.6 ± 161

Nitrogen Fixation (as ARA) and Nodule Biomass

Rates of N fixation as ARA (Table 3) were generally higher in the first year of the study (35-198 $\mu\text{mol g}^{-1} \text{h}^{-1}$) than the second (17-37 $\mu\text{mol g}^{-1} \text{h}^{-1}$), and were also higher than many reported values of ARA for *A. rubra*. A range of reported values in the literature was 3.2-90 $\mu\text{mol g}^{-1} \text{h}^{-1}$ (Binkley, 1981; Monaco et al., 1982; Rojas et al., 2001; Teklehaimanot & Martin, 1999). The conditions of these studies were more reflective of native growing conditions, *Alnus* sp. are commonly

found in poor-quality, low-productivity soils. The high amount of biosolids that were used in this study created unusual growing conditions. The first year ARA for compost (108 \pm 52 $\mu\text{mol g}^{-1} \text{h}^{-1}$) and high metal (105 \pm 100 $\mu\text{mol g}^{-1} \text{h}^{-1}$) treatments were relatively high as compared to the controls (35 \pm 16 $\mu\text{mol g}^{-1} \text{h}^{-1}$). However, as a consequence of high variability, this increase in ARA can only be viewed as a trend ($P = 0.05$). Similar variability has also been seen with other studies of *Alnus* sp. (Beaupied et al., 1990; Rytter et

Table 2 - Mean and standard deviation for soil NO_3^- , NH_4^+ , total C, N, C:N ratio and pH by treatment type and amount, both years of the study.

Year 1							
Treatment	Rate	NO_3^-	NH_4^+	C	N	C:N	pH
	Mg ha^{-1}	----- mg kg^{-1} -----		----- g kg^{-1} -----			
Control	0	30 \pm 18*	8 \pm 4*	1 \pm 1 a	0.1 \pm 0.1 a	10 \pm 2 a	5.1 \pm 0.3 b
Compost	250	109 \pm 95*	37 \pm 8*	19 \pm 8 c	0.6 \pm 0.2 b	30 \pm 1 c	5.3 \pm 0.1 bc
Compost	500	47 \pm 2*	61 \pm 16*	31 \pm 5 d	1.2 \pm 0.1 c	26 \pm 2 b	5.5 \pm 0.1 c
Compost	1000	221 \pm 229*	48 \pm 14*	35 \pm 6 d	1.2 \pm 0.3 c	29 \pm 2 c	5.3 \pm 0.3 bc
High metal	250	98 \pm 20*	46 \pm 17*	4 \pm 2 ab	0.4 \pm 0.2 ab	10 \pm 1 a	4.6 \pm 0.2 a
High metal	500	222 \pm 98*	110 \pm 43*	7 \pm 1 ab	0.7 \pm 0.1 b	9 \pm 1 a	4.7 \pm 0.1 a
High metal	1000	159 \pm 35*	61 \pm 41*	12 \pm 3 bc	1.4 \pm 0.3 c	9 \pm 0 a	4.6 \pm 0.1 a
Year 2							
Treatment	Rate	NO_3^-	NH_4^+	C	N	C:N	pH
	Mg ha^{-1}	----- mg kg^{-1} -----		----- g kg^{-1} -----			
Control	0	9 \pm 8 a	5 \pm 6 a	2 \pm 1 a	0.3 \pm 0.2 a	7 \pm 1 a	5.1 \pm 0.5 c
Compost	250	29 \pm 28 a	29 \pm 15 a	10 \pm 4 ab	0.4 \pm 0.2 a	22 \pm 2 bc	4.9 \pm 0.1 bc
Compost	500	30 \pm 26 a	76 \pm 62 a	18 \pm 7 bc	0.9 \pm 0.2 bc	20 \pm 3 b	5.1 \pm 0.1 c
Compost	1000	134 \pm 40 b	192 \pm 135 b	23 \pm 8 c	1.0 \pm 0.4 c	24 \pm 2 c	5.0 \pm 0.1 bc
High metal	250	19 \pm 7 a	14 \pm 5 a	4 \pm 1 a	0.5 \pm 0.0 ab	8 \pm 1 a	4.6 \pm 0.2 b
High metal	500	50 \pm 13 a	26 \pm 23 a	6 \pm 3 a	0.6 \pm 0.2 abc	9 \pm 2 a	4.6 \pm 0.2 b
High metal	1000	139 \pm 15 b	77 \pm 30 a	9 \pm 2 a	0.9 \pm 0.2 c	9 \pm 1 a	4.1 \pm 0.1 a

Different letters following numbers indicate significantly different values using the Duncan post-hoc mean separation procedure ($P = 0.05$), values followed by *are not statistically different

Table 3 - Mean and standard deviation for acetylene reduction activity (ARA) and nodule biomass for both years of the study. Total nodule biomass was not measured in year two.

Treatment	Rate	Year 1		Year 2	
		ARA		ARA	
		$\mu\text{mol C}_2\text{H}_4 \text{g}^{-1} \text{hr}^{-1}$		$\mu\text{mol C}_2\text{H}_4 \text{g}^{-1} \text{hr}^{-1}$	
	Mg ha^{-1}		Nodule Biomass		
		$\mu\text{mol C}_2\text{H}_4 \text{g}^{-1} \text{hr}^{-1}$	g		$\mu\text{mol C}_2\text{H}_4 \text{g}^{-1} \text{hr}^{-1}$
Control	0	35 \pm 16	0.23 \pm 0.17		23 \pm 14
Compost	250	105 \pm 69	0.19 \pm 0.07		37 \pm 23
Compost	500	99 \pm 64	0.21 \pm 0.10		19 \pm 8
Compost	1000	100 \pm 47	0.60 \pm 0.29		19 \pm 5
High metal	250	51 \pm 57	0.37 \pm 0.28		29 \pm 36
High metal	500	198 \pm 103	0.37 \pm 0.14		33 \pm 30
High metal	1000	74 \pm 95	0.21 \pm 0.07		17 \pm 25

al., 1991; Verghese & Misra, 2000), and is to be expected as *Alnus* and *Frankia* are “wild type” organisms and are more heterogeneous than agronomic Rhizobium and legume species. Nodule biomass (Table 3) was also not statistically different between treatments (only collected first year) and no relation to ARA was seen ($P = 0.05$).

Decreased nodulation and suppression of N fixation as a result of soil available N has been well documented for *Alnus* (Huss-Danell et al., 2001; Martin et al., 2003; Wall, 2000), but in this study the reverse was observed. At the 500 Mg ha⁻¹ amendment the high metal biosolids amendment added 12 Mg total N ha⁻¹ and the compost 9 Mg total N ha⁻¹. The portion of total N in the amendments that is plant available over a growing season will vary; however, in all likelihood the amount of added N that amendments supplied was more than the plants were able to use for a single season. In a study by Baker et al. (1997), although the addition of NO₃⁻ to *A. glutinosa* was found to have an inhibitory affect on ARA, addition of NH₄⁺ did not. NH₄⁺ was found to reduce nodule size, but not rates of ARA. A similar response with *A. rubra* was seen in this study, trees grown in biosolid amended soils had higher rates of ARA than trees grown in soils that had been amended with an equivalent amount of urea. This increase in ARA may be related to the P in the amendments as increases in nodulation and N fixation with P additions have also been well documented (Gentili & Huss-Danell, 2003; Uliassi & Ruess, 2002). Prior to installation of the study, total P in high metal and compost amendments were 13.3 and 8.0 g kg⁻¹ respectively. However, P also does not solely explain the high rates of ARA seen in the first year of this study, as the rates of ARA were all lower the second year when P concentrations would still be relatively high. In addition, background soil total P concentrations were 1.54 g kg⁻¹, which is relatively high (Brady & Weil, 2002).

These results suggest that the *Alnus-Frankia* relationship in this study was not negatively impacted by the high metal treatments. Metals resistance of *Frankia* has been studied in controlled environments with metals added as salts. Sayed et al. (2000) grew 3 strains of *Frankia* in culture with Pb, cobalt (Co) and aluminum (Al) added as salts and measured their relative growth rates and found tolerance to metals varied with strain. Richards et al. (2002) grew twelve strains of *Frankia* on plates or in broth with metals added as salts (AgNO₃, Na₂HAsO₄, NaAsO₂, CdCl₂, CoCl₂, K₂CrO₄, CuCl₂, NiCl₂, Pb(NO₂)₂, K(SbO)C₄H₄O₆ and Na₂SeO₂) and found some strains showed resistance to heavy metals, but again, the results varied among strains. As this study was conducted with wild

type *Frankia* and profiling of the organisms was beyond the scope of this research, it is not possible to draw conclusions on the response of specific strains present at the field site.

Shoot Biomass

At the time of the first years tree harvest there were no differences in biomass, but the second year biomass did differ by treatment type ($P = 0.05$), (Figure 1). The control trees had the smallest shoot biomass (120 ± 66 g), followed by trees in the high metal treatments (327 ± 78 g) with the highest yield in the compost treatments (544 ± 200 g). In compost treatments, there was a moderate trend of increased biomass with increasing amount of amendment ($r^2 = 0.36$). In the high metal treatments, the trend appeared to be decreasing biomass with increasing application rates.

It is difficult to determine an exact cause for the biomass of trees in the high metal treatments being less than the compost treatments. Although metals have been seen to have an effect on shoot biomass (El-Kenawy et al., 1997), they were not a clear contributor in this study. No correlation was seen for soil metal concentrations and shoot biomass, and shoot biomass was consistently higher in high metal treatments than in the control. This was the case even though soils in these treatments also had lower pH which results in higher metal bioavailability. Increased yield observed in this study may be the result of the added nutrients in the biosolids outweighing the nega-

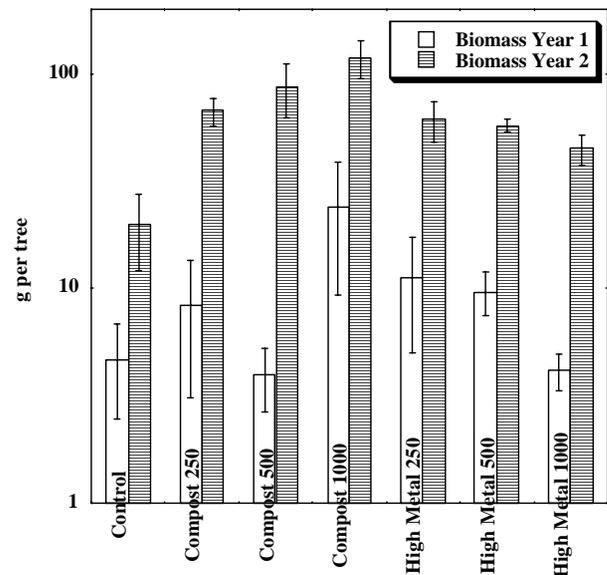


Figure 1 - Shoot biomass results for both years by treatment type and rate. Numbers on bars show average biomass in grams of dry weight for each treatment type. Data, means and standard deviation are shown for two years.

tive impact of metals. However, *Alnus* have also exhibited metal tolerance when grown on metal contaminated sites. Kramer et al. grew *A. viridis* (Sitka alder) in a greenhouse study (Kramer et al., 2000b) and a field study (Kramer et al., 2000a) in biosolids amended copper (Cu) mine tailings and found the greatest uptake of nutrients, largest biomass, and smallest root:shoot in the biosolids amended mine tailings over unamended tailings, as well as uncontaminated control soils. Total soil metal concentrations were not reported, but water quality criteria at the site is exceeded for Cd, Cu and Zn during spring snowmelt, with Cu being the highest concern (Forest_Service, 2004). *Frankia* in culture has also shown metal tolerance, as discussed above. Foliar uptake of metals as a metric of bioavailability also did not show a negative impact and is discussed next.

Foliar Metals

Foliar Cd and Pb were below detection in all treatments, foliar Zn results are presented in Figure 2. Detection limits for Cd, Pb and Zn were 0.6, 1.2 and 0.6 mg kg⁻¹ respectively. Although there was a weak relationship between increasing amount of amendment and increased foliar Zn ($r^2 = 0.29$), there was no difference in foliar Zn as a function of type of biosolid amendments. Total soil Zn in the 1000 Mg ha⁻¹ compost treatment was 279 mg kg⁻¹, with a corresponding foliar Zn of 249 mg kg⁻¹; total soil Zn in the 1000 Mg ha⁻¹ high metal treatment was 1760 mg kg⁻¹, with a foliar Zn of 274 mg kg⁻¹. Based on biomass results, visual appearance and comparison to values in the lit-

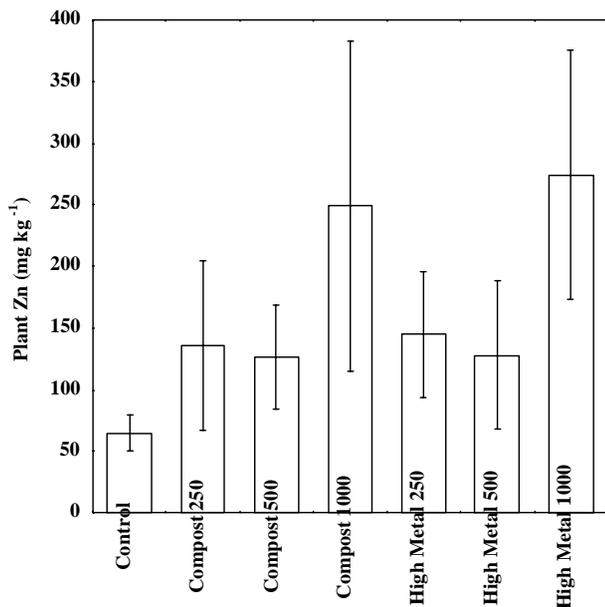


Figure 2 - Average values for foliar Zn by treatment type and rate for the second year of the study. Data, means and standard deviation are shown.

erature, plants in all treatments showed no evidence of phytotoxicity. Values for foliar metals in *A. rubra* growing under similar conditions were not found in the literature, but foliar metals were low as compared to reported values for other *Alnus* sp. growing in high metal soils with biosolid amendments. For example, Sopper (1989) found foliar Cd, Pb and Zn of *A. glutinosa* to be 3-5, 76-140, and 807-2070 mg kg⁻¹ respectively at a highly contaminated Zn smelter site amended with biosolids and ash. Surface (0-30 cm) soil concentrations were: 31,316 mg Zn kg⁻¹, 5,225 mg Pb kg⁻¹, and 1,248 mg Cd kg⁻¹. Kramer et al. (2000a) saw foliar Zn of 546 mg kg⁻¹ without growth being negatively impacted. A wide variation in uptake patterns is seen even within genotypes of the same species, so it is unknown exactly how these values compare to *A. rubra*, but if comparison is valid it appears that metals in the high metal treatments were mostly not in a bioavailable form. The factors contributing to low metal availability in this study are not clear, although biosolids used as soil amendments have been shown to increase the sorption of heavy metals (Li et al., 2001), and reduce the bioavailability of metals (Brown et al., 1998; DeVolder et al., 2003). Soil organic matter and/or N could possibly have been a factor in increasing foliar Zn with increasing amount of both types of biosolid amendments, Hutchinson et al. (1998) found N added as (NH₄)₂SO₄ to increase foliar Cd, and Zn in tree rings, of *Acer saccharum* (sugar maple), with foliar Cd in high and medium treatments being two times the concentrations in the low and control treatments. Untreated soil metals concentrations were 0.06-0.08 and 50-64 mg kg⁻¹ for Cd and Zn with foliar concentrations of 0.09-0.10 and 10-25 mg kg⁻¹ for Cd and Zn. In this study foliar Zn correlated to concentrations of available N ($r^2 = 0.42$), and total soil N ($r^2 = 0.44$) more strongly than the correlation to soil Zn concentrations ($r^2 = 0.26$). Soil pH was not a predictor of foliar Zn.

CONCLUSION

This study focused on potential impacts to the *Alnus/Frankia* relationship as a result of elevated soil metals resulting from the land application of biosolids. For this study, increased growth as a result of biosolids application, without an impact on N fixation, was the primary response observed for trees grown in both low and high metal concentration biosolids. These results illustrate the potential benefits associated with biosolid application. Although these benefits were observed for both materials, this should not be taken as a justification for permitting land application of high metal biosolids. The metal concentrations of the historic

biosolids are well above current regulatory limits in the US and would not be suitable for land application. Although tree growth in this instance was enhanced, the potential for negative ecosystem effects as a result of use of these materials was not fully considered in this study. The efficacy of pre-treatment regulations can be seen in the dramatic decrease in total metal concentrations from the historic material and the current day biosolids, both from the same wastewater treatment plant. The full benefits of these materials can be realized if every reasonable attempt is made to reduce contaminant concentrations.

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